

Parent, Alkylated, and Sulfur/Oxygen-Containing Polycyclic Aromatic Hydrocarbons in Mainstream Smoke from 13 Brands of Chinese Cigarettes

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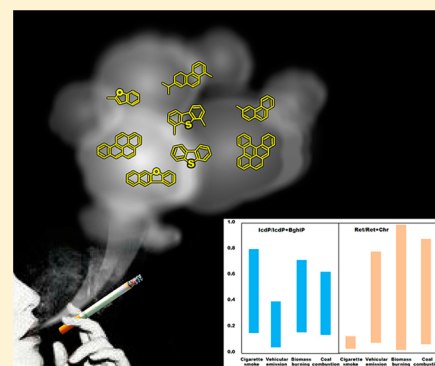
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Supporting Information

ABSTRACT: China has the world's largest population of smokers with serious health consequences, yet we know a very limited spectrum of hazardous chemicals in cigarette smoke even for carcinogenic polycyclic aromatic hydrocarbons (PAHs). Here, we chose 13 popular cigarette brands sold in China markets, collected particulate matters in mainstream smoke using filter pads and an automatic smoking machine, and analyzed 56 PAHs, including 31 parent, 18 alkylated, and 7 sulfur/oxygen-containing PAHs (S/O PAHs). The 56 PAHs in mainstream smoke totaled from 244.2 ± 28.5 to 10254.8 ± 481.5 ng cig⁻¹; parent, alkylated, and S/O PAHs shared 16–23%, 64–74%, and 6–18%, respectively. Benzo[a]pyrene (BaP) ranged 1.1–41.6 ng cig⁻¹, while BaP equivalent concentrations (BaP_{eq}) ranged 3.6–120.2 ng cig⁻¹, but contributions to BaP_{eq} by individual carcinogenic PAH species varied with cigarette brands. When these cigarette smoke source profiles were pooled together with those of other combustion ones available in the literature, we found that widely used diagnostic ratios of parent PAHs failed to distinguish cigarette smoke from other combustion sources, except that the ratio indeno[1,2,3-cd]pyrene/(indeno[1,2,3-cd]pyrene + benzo[g,h,i]perylene) can largely separate cigarette smoke from vehicular emissions and that the ratio of Retene/(Retene + chrysene) can further discriminate cigarette smoke from coal combustion when alkylated PAHs are involved.



INTRODUCTION

Polycyclic aromatic hydrocarbons (PAHs) are known to cause cancer in humans. They are ubiquitous in ambient air largely from incomplete combustion of fossil fuels and/or biomass burning, yet environmental tobacco smoke (ETS), either mainstream or side-stream, contributes substantially to PAHs in indoor microenvironments,^{1,2} particularly in smoker's homes.³ As PAHs are among the most potent carcinogens in cigarette smoke and tobacco smoke-derived PAHs bear a large burden in human exposure to this class of carcinogens,⁴ characterizing compositions and levels of PAHs in cigarette smoke is of wide and enduring concern for the protection of public health.

Mainstream smoke (MS), referring specifically to the smoke that a smoker inhales and then exhales, represents the direct health hazard to smokers though voluntary inhalation. Previous studies have revealed the presence of PAHs and other pollutants in MS,^{1,5–7} and among them, benzo[a]pyrene

(BaP) was the most studied species due to its ability to induce lung tumors.⁸ Besides BaP, some other PAHs within the USEPA's 16 priority PAHs are also identified as probably or possibly carcinogenic to humans by agencies like the International Agency for Research on Cancer (IARC).¹ Consequently, the 16 priority PAHs or their subsets in ETS were investigated concerning their levels, compositions,^{1,6,9} and toxicity,⁵ as well as the influence of smoking conditions and cigarette design on the chemical delivery.^{7,10,11} However, as cigarette smoke is a complex mixture with multiple classes of chemical compounds, more PAHs, such as alkylated and sulfur/oxygen (S/O) PAHs, also exist in cigarette smoke,^{12,13} either being possible human carcinogens or promoters for the carcinogenicity of other

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PAHs.⁶ Previous studies have evidenced that some alkylated and S/O PAHs were more toxic than their parent ones.^{14,15} Hence, accurate evaluation of the toxicity of cigarette smoke requires more PAH species to be measured. Unfortunately, there is little quantitative data about these additional PAHs. Only a few studies reported the concentrations of methyl-naphthalenes,⁷ methylphenanthrenes/methylanthracenes and methylpyrenes,¹² and 5-methylchrysene¹⁶ in the MS particulates.

China is the world's leading tobacco producer and has a third of the world's tobacco consumers.⁴ However, most of the published researches focused on PAHs in reference cigarettes and/or commercial cigarette brands outside of China. Only a few studies are available for cigarettes smoke PAHs in China,^{4,17,18} and the target compounds were still limited within the 16 priority PAHs. To our best knowledge, only one study published in English reported 9 priority PAHs and other harmful chemicals in 20 brands of Chinese cigarettes.⁴ In this study, we measured the levels of 56 PAHs, including 31 parent, 18 alkylated, and 7 S/O PAHs in the MS particulates from 13 popular Chinese cigarette brands so as to complement these deficiencies. PAH levels in reference cigarettes were also measured for comparison with previous studies. The BaP equivalent (BaP_{eq}) concentrations were further calculated to indicate carcinogenic potency of PAHs using our current data. As diagnostic ratios have been widely used to identify PAH sources^{19,20} and utilizing the alkylated and S/O analogues may facilitate source discrimination,^{21,22} here, we finally evaluate whether PAH (including parent, alkylated, and S/O PAHs) ratios can distinguish cigarette smoke from other sources, using published source profiles for combustion sources together with our data.

EXPERIMENTAL SECTION

Materials. Sixteen priority PAHs and deuterated PAHs (naphthalene-d8, acenaphthene-d10, phenanthrene-d10, chrysene-d12, perylene-d12) were obtained from Ultra Scientific, Inc. (North Kingstown, RI, U.S.A.). A standard mixture solution of 10 PAHs (1-methylnaphthalene, 2-methylnaphthalene, biphenyl, 2,6-dimethylnaphthalene, 2,3,6-trimethylnaphthalene, 2-methylphenanthrene, 1-methylphenanthrene, 3,6-dimethylphenanthrene, benzo[e]pyrene and perylene) was obtained from AccuStandards, Inc. (New Haven, CT, U.S.A.). A standard mixture solution of 8 S/O PAHs (dibenzofuran, 1-methyldibenzofuran, dibenzothiophene, 4-methyldibenzothiophene, 2-methyldibenzothiophene, 3-methyldibenzothiophene, benzonaphtho[2,1-d]thiophene, benzonaphtho[2,3-d]thiophene) was purchased from Chiron AS (Trondheim, ST, Norway). Hexamethylbenzene and coronene were purchased from Aldrich Chemical Co. (Milwaukee, WI, U.S.A.). Reference cigarettes, 1R3F and 2R4F, were obtained from the University of Kentucky (Lexington, KY, U.S.A.). Domestic commercial cigarettes were purchased from various retail sources in China. Cambridge filter pads (CFPs) used to collect MS particulate matter were obtained from Whatman (Maidstone, U.K.).

Smoke Collection. Prior to smoking, the cigarettes and CFPs were conditioned at constant relative humidity (60%) and temperature (22 °C) for no less than 48 h using a temperature and humidity chamber (WTC binder, Germany). MS total particulate matter (TPM) generated following U.S. Federal Trade Commission (FTC)/International Standard Organization (ISO) conditions (60 s puff interval, 2 s puff duration, and 35 mL puff volume) was collected on individual CFPs using a Borgwaldt (Germany) RM200 20-port automatic

smoking machine. The cigarettes were smoked to a butt length of 23 mm or the length of the filter overwrap plus 3 mm, whichever was longer, using the industry-standard Cambridge filter pad holder. Three replicate samples were collected per cigarette type with 20 cigarettes smoked for each sample. Table S1 of the Supporting Information presents the specification and average puff count for each type of cigarette.

Sample Preparation and GC/MS Analysis. The analytical procedure used for extraction, separation, and measurement of PAHs was detailed elsewhere,^{23,24} and only a brief description is given here. Before extraction, each CFP was spiked with the deuterated PAH standards as surrogates, and then Soxhlet-extracted with dichloromethane for 72 h. The extracts were concentrated to 1 mL, solvent-exchanged to redistilled hexane, and then separated into two fractions using a 1:2 alumina/silica column chromatography. The first fraction containing nonpolar compounds was eluted by 30 mL of hexane. The second fraction, containing PAHs, was eluted by 70 mL of dichloromethane/hexane (3:7 v/v), concentrated to ~1 mL, and blown to ~200 μ L under a gentle stream of nitrogen. A total of 8 μ L of a 50 μ g mL⁻¹ hexamethylbenzene internal standard was added to the samples before instrumental analysis.

The samples were analyzed using an Agilent GC-MS (6890–5973N) equipped with a HP-5 capillary column (50 m \times 0.32 mm \times 0.17 μ m). The column temperature was initiated at 80 °C (held for 5 min) and increased to 290 °C at 3 °C min⁻¹ (held for 30 min). An aliquot of 1 μ L was injected in split mode with a solvent delay of 6 min. Parent, alkylated, and S/O PAHs were identified based on their mass spectra and retention times. For those with authentic standards, quantitation was performed using the internal calibration method with the five-point calibration curves, and for alkylated PAHs without authentic standards, their amounts were approximated with their isomers/homologues or parent PAHs closest in retention times as the alternative standards.

Recoveries were evaluated by the deuterated surrogates spiked to the samples. The mean recoveries were 76 \pm 20%, 89 \pm 13%, 98 \pm 15%, 89 \pm 16%, and 82 \pm 16%, respectively, for naphthalene-d8, acenaphthalene-d8, phenanthrene-d10, chrysene-d12, and perylene-d12. The relative standard deviations (RSDs) of individual PAHs except for naphthalene were all within 15% among the three replicate samples collected per cigarette brand.

RESULTS AND DISCUSSION

PAH Levels in Reference Cigarettes. We measured 56 PAHs in MS for 1R3F and 2R4F Kentucky reference cigarettes using our method (Table S2, Supporting Information). Individual PAHs for 1R3F cigarettes showed levels 1.1–2.5 times that for 2R4F ones, and the total PAH levels in 1R3F (4208.8 \pm 746.7 ng cig⁻¹) were 1.6 times that in 2R4F (2564.3 \pm 134.8 ng cig⁻¹). However, TPM normalized PAH levels (PAHs/TPM) for 1R3F (303.9 \pm 51.6 ng mg⁻¹) were just slightly higher than that for 2R4F (272.9 \pm 5.3 ng mg⁻¹). Considering 1R3F is close to 2R4F in tobacco blend, higher PAH yields for 1R3F may be due to its lower filter ventilation than 2R4F. 2R4F is designed as the substitute of 1R4F,²⁵ and both of them are examples of American blended cigarettes with filler consisting of bright, burley, oriental, and reconstituted tobaccos.¹ Literature data of these two types of reference cigarettes smoked under FTC/ISO conditions are listed in Table 1 for comparison. Among the PAH levels for 2R4F

Table 1. Measured PAH Levels (ng cig⁻¹) in Reference Cigarettes from Different Studies

| | 1R3F ^a | 2R4F ^a | 2R4F ^b | 2R4F ^c | 2R4F ^d | 1R4F ^d | 1R4F ^e | 1R4F ^f | 1R4F ^g | 1R4F ^h |
|--|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|-------------------|
| | GC-MS | GC-MS | LC-MS/MS | GC-MS | GC-MS | GC-MS | GC-MS | GC-MS | GC-MS | GC-MS |
| Nap ⁱ | 90.5 | 36.8 | | | 192 | 350.3 | 236 | | 292.8 | 361.7 |
| Acey | 45.5 | 26.1 | | | 88.3 | 116.9 | 50.4 | | | |
| Ace | 31.4 | 19.7 | | | 51.3 | 84.8 | 25.3 | | | |
| Fl | 168.9 | 103.8 | | | 156.3 | 217.5 | 119 | | 144.5 | 239 |
| Phe | 126.4 | 90.5 | | | 145.2 | 134.8 | 110 | | 76.3 | 147.7 |
| Ant | 61.7 | 41.8 | | | 69.8 | 74.9 | 38.1 | | 34.5 | 35.8 |
| Flu | 72.5 | 46.6 | | | 63 | 74.4 | 46.2 | | 42.6 | 51.6 |
| Pyr | 53.4 | 38.7 | | 35.3 | 49.5 | 48.6 | 33.2 | | 27.9 | 32.1 |
| BaA | 22.2 | 18.1 | 33.5 | 11.2 | 16.5 | 13.4 | 13.2 | 9.4 | 10.5 | 14 |
| Chr | 24.7 | 17.7 | | | 19.7 | 15.7 | 21.8 | | 14 | 11.2 |
| BbF | 8.2 | 5.4 | 5.9 | 4.52 | 10.6 | 9.4 | 8.6 | 4.8 | | 11.2 ^j |
| BkF | 7.0 | 5.0 | 2.2 | 1.74 | 1.9 | 1.5 | 1.5 | 1.4 | | |
| BeP | 13.9 | 8.6 | | | 3.2 | 2.9 | 4 | 6 | 4 | 6.4 |
| BaP | 18.6 | 13.4 | 8.8 | 6.21 | 11 | 10.3 | 7.9 | | 4.6 | 7.6 |
| Pyl | 2.1 | 1.7 | | | | | | | 0.9 | 3.5 |
| IcdP | 8.2 | 5.2 | 5.9 | 2.55 | | | 3.5 | 2.5 | | |
| DahA | 1.1 | 0.8 | 4.7 | <0.97 | | | 0.6 | 0.7 | 0.4 | 2.3 |
| BghiP | 2.1 | 1.4 | | | | | 2.5 | | 1 | 2.3 |
| TPM ^k (mg cig ⁻¹) | 13.8 | 9.4 | | 9.3 | | | 11 | 10.3 | 10.9 | |

^aThis study. ^bRef 16. ^cRef 26. ^dRef 1. ^eRef 27. ^fRef 5. ^gRef 28. ^hRef 29. ⁱCompounds are given abbreviations as follows: naphthalene (Nap), acenaphthylene (Acey), acenaphthene (Ace), fluorene (Fl), phenanthrene (Phe), anthracene (Ant), fluoranthene (Flu), pyrene (Pyr), benz[a]anthracene (BaA), chrysene (Chr), benzo[b]fluoranthene (BbF), benzo[k]fluoranthene (BkF), benzo[e]pyrene (BeP), benzo[a]pyrene (BaP), Perylene (Pyl), indeno[1,2,3-cd]pyrene (IcdP), dibenz[a,h]anthracene (DahA), and benzo[g,h,i]perylene (BghiP). ^jBbF+BkF. ^kTotal particulate matter.

determined by previous studies,^{1,16,26} the TPM level was only reported by Roemer et al.,²⁶ and it was almost the same as that measured in our study (Table 1), although the PAHs emitted per cigarette in our study were much higher. Compared to the amounts of PAHs emitted per cigarette by Ding et al.,¹ our results were lower for most PAHs except for BaA, BkF, BeP, and BaP. However, levels of BbF and IcdP quite approximated that reported by Ding et al.¹⁶ The differences in measured PAH levels for 2R4F between studies might be partly caused by the numbers of cigarettes being smoked; only 1–3 cigarettes were smoked by Ding et al.,^{1,16} whereas 20 or 30 cigarettes were smoked by Roemer et al.²⁶ and 20 cigarettes were smoked in this study. Different pretreatment and instrumental determination procedures could also be reasons for the different levels measured for the same species. As shown in Table 1, none of previous studies reported all the listed 18 PAH species for 2R4F or 1R4F as we did. Comparison of our 2R4F results with those reported for 1R4F also revealed similar TPM levels, while the individual PAHs fell into three categories; the levels of Nap, Acey, Ace, and Fl were lower in 2R4F than in 1R4F; the levels of BaA, BkF, BeP, BaP; and IcdP in 2R4F were higher in 2R4F than in 1R4F; and the levels of remaining PAHs in 2R4F fell into the range of literature data for 1R4F (Table 1).

PAH Levels in Different Commercial Cigarettes. Levels of individual PAHs in the MS of 13 commercial cigarettes are listed in Table S2 of the Supporting Information in comparison with previous particle-phase data for various cigarettes smoked using FTC/ISO conditions.^{1,4–7,9–11,16,27,28} The ranges for 6 parent PAHs, namely, BaA, BbF, BkF, BaP, DahA, and IcdP, were wider than those reported in the IARC monograph,³⁰ in which the data were obtained more than 30 years ago.

Emission of individual PAH compounds varied from 0.1 to 1123.1 ng cig⁻¹ and also exhibited large variations across different brands, despite that they fell into the broad range of

literature data, which covered 1–3 orders of magnitude (Table S2, Supporting Information). The average RSD for individual species ranged from 21% for dibenzothiophene (DBT) to 182% for benzo[c]phenanthrene, with an average of 52%. The variations in the compositions of PAHs could be associated with many factors including tobacco types and blends; ingredient types; filter types, size and ventilation; manufacturing processes; and even analytical procedures.¹ Notably, PAHs generated from Kent were 1 order of magnitude less than those from others probably due to higher ventilation.

The average TPM yield for the cigarettes tested was 15.7 ± 4.7 mg cig⁻¹, with a less variation (RSD~30%) compared to PAHs. These TPM levels were similar to that of 15.5 ± 3.0 mg cig⁻¹ reported by Akpan et al.⁴ for 20 brands of Chinese cigarettes and close to those for some custom-made blended cigarettes (8.61–11.01 mg cig⁻¹)¹⁰ and blended United States market cigarettes (2–28.6 mg cig⁻¹),⁵ but much lower than the mean value determined by Moir et al. (46.9 mg cig⁻¹).⁷ A Pearson correlation analysis indicated most PAHs were strongly correlated with each other ($r = 0.514–0.996$, $p < 0.05$) and with TPM ($r = 0.522–0.862$, $p < 0.05$). The average RSD of PAHs/TPM was slightly lowered (RSD 47%) compared to that of PAHs.

Table 2 presents a summary of total and grouped PAHs from this study. Total concentrations of PAHs ranged from 244.2 ± 28.5 ng cig⁻¹ in Kent to 10254.8 ± 481.5 ng cig⁻¹ in YZB. Among the three groups of PAHs, alkylated PAHs were the most abundant and comprised of 64–74% of the total PAH concentrations, followed by parent PAHs (16–23%) and S/O PAHs (6–18%). If we further divided the parent PAHs into three groups based on their molecular weights (MW) as low MW parent PAHs (LP-PAHs, MW 128–178 amu), medium MW parent PAHs (MP-PAHs, MW 202–228 amu), and high MW parent PAHs (HP-PAHs, MW 252–302 amu), LP-PAHs

Table 2. Total Concentrations (ng cig⁻¹, average ± standard deviation) of PAH Groups and BaP_{eq} in Commercial Cigarettes

| brand | LP-PAHs ^a | MP-PAHs ^b | HP-PAHs ^c | ∑PPAHs ^d | ∑APAHs ^e | ∑SOPAHs ^f | ∑PAHs | ∑ ₁₆ PAHs ^g | BaP _{eq} ^h | TPM (mg cig ⁻¹) |
|-------------------|----------------------|----------------------|----------------------|---------------------|---------------------|----------------------|-----------------|-----------------------------------|--------------------------------|-----------------------------|
| Blended | | | | | | | | | | |
| Kent | 18.4 ± 4.7 | 16.2 ± 1.8 | 5.7 ± 0.4 | 40.3 ± 6.8 | 159.1 ± 17.2 | 44.8 ± 7.0 | 244.2 ± 28.5 | 34.3 ± 6.0 | 3.6 ± 0.9 | 7.1 ± 0.5 |
| WBL | 553.7 ± 49.0 | 168.2 ± 5.3 | 84.8 ± 11.5 | 806.7 ± 49.4 | 2894.8 ± 173.1 | 228.4 ± 29.6 | 3930.0 ± 232.2 | 688.0 ± 47.9 | 66.1 ± 23.4 | 11.8 ± 0.3 |
| 555 | 565.8 ± 23.8 | 193.1 ± 0.5 | 76.8 ± 2.8 | 835.6 ± 20.7 | 2357.0 ± 87.1 | 481.2 ± 22.9 | 3673.8 ± 129.4 | 716.0 ± 19.3 | 52.5 ± 1.9 | 8.8 ± 0.2 |
| YC | 582.4 ± 62.7 | 171.2 ± 20.9 | 60.4 ± 4.7 | 814.0 ± 87.9 | 2358.3 ± 269.6 | 407.6 ± 61.2 | 3580.0 ± 415.9 | 688.9 ± 75.4 | 38.4 ± 7.8 | 13.7 ± 0.1 |
| Flue-cured | | | | | | | | | | |
| HTS | 937.3 ± 41.1 | 362.2 ± 5.2 | 148.4 ± 3.7 | 1447.9 ± 36.8 | 4588.9 ± 66.1 | 635.8 ± 18.2 | 6672.6 ± 44.7 | 1254.3 ± 31.4 | 99.9 ± 5.2 | 14.0 ± 0.3 |
| HGS | 1196.2 ± 127.8 | 335.4 ± 17.1 | 129.7 ± 14.2 | 1661.3 ± 146.9 | 5043.5 ± 266.0 | 923.1 ± 7.4 | 7628.0 ± 403.1 | 1414.8 ± 126.6 | 96.7 ± 8.8 | 15.2 ± 0.4 |
| JZ | 1175.3 ± 50.6 | 394.0 ± 10.1 | 166.7 ± 9.0 | 1736.0 ± 66.5 | 5289.5 ± 208.5 | 859.0 ± 57.4 | 7884.4 ± 316.9 | 1477.7 ± 60.6 | 120.2 ± 8.6 | 17.5 ± 0.3 |
| LQ | 930.6 ± 856.9 | 260.6 ± 222.3 | 160.2 ± 82.6 | 1351.4 ± 1147.8 | 4130.4 ± 3635.9 | 615.9 ± 531.9 | 6097.7 ± 5310.2 | 1134.2 ± 982.2 | 82.7 ± 36.0 | 21.0 ± 0.9 |
| YY | 998.3 ± 142.2 | 375.5 ± 46.7 | 150.4 ± 21.0 | 1524.1 ± 207.2 | 4894.0 ± 914.7 | 820.9 ± 119.7 | 7239.1 ± 1233.1 | 1269.1 ± 159.8 | 108.3 ± 12.9 | 13.9 ± 1.3 |
| YZB | 1559.5 ± 44.9 | 381.9 ± 86.9 | 164.1 ± 22.3 | 2105.4 ± 113.5 | 7063.4 ± 399.6 | 1085.9 ± 7.0 | 10254.8 ± 481.5 | 1775.7 ± 99.9 | 115.3 ± 26.2 | 22.0 ± 0.7 |
| BS | 586.1 ± 23.0 | 224.9 ± 6.2 | 89.4 ± 0.1 | 900.4 ± 17.9 | 2627.8 ± 85.5 | 435.4 ± 24.5 | 3963.7 ± 108.2 | 773.0 ± 13.0 | 60.8 ± 0.1 | 14.8 ± 0.1 |
| ZH | 669.8 ± 73.7 | 256.7 ± 11.6 | 106.8 ± 3.8 | 1033.2 ± 72.0 | 3222.0 ± 225.4 | 533.9 ± 8.4 | 4789.2 ± 299.1 | 883.1 ± 60.2 | 70.0 ± 2.9 | 18.1 ± 0.1 |
| WYS | 931.2 ± 94.7 | 265.0 ± 23.8 | 108.1 ± 7.8 | 1304.3 ± 115.4 | 3895.1 ± 284.4 | 607.0 ± 27.0 | 5806.4 ± 413.5 | 1090.4 ± 100.9 | 72.8 ± 9.0 | 18.5 ± 0.2 |

^aMeasured parent PAHs with molecular weight 128–178. ^bMeasured parent PAHs with molecular weight 202–228. ^cMeasured parent PAHs with molecular weight 252–302. ^dTotal measured parent PAHs. ^eTotal measured S/O PAHs. ^fTotal measured alkylated PAHs. ^gTotal 16 USEPA priority PAHs. ^hDibenzopyrenes were not included in the calculation.

dominated the total parent PAHs, with shares ranging from 46% to 74%, followed by MP-PAHs (18–40%) and HP-PAHs (7–14%). This dominance of LP-PAHs over MP- and HP-PAHs in MS of cigarettes was also observed in many previous studies.^{1,6,7,11} At low to moderate combustion temperature, alkylated and LP-PAHs are abundant,²⁴ consistent with the feature of cigarette smoking.⁵

The U.S. Environmental Protection Agency (USEPA) has identified 16 priority PAHs on the basis of evidence that they cause or may cause cancer. Total 16 priority PAH concentrations ($\sum_{16}\text{PAHs}$) in the 13 commercial cigarettes ranged 34.3–1775.7 ng cig⁻¹ (Table 2), close to that in other cigarettes reported by Hearn et al. (241.7–471.8 ng cig⁻¹, without IcdP, DahA and BghiP),¹¹ Ding et al. (86.2–2690.7 ng cig⁻¹, without IcdP, DahA and BghiP),¹ Ding et al. (797.6–2664.8 ng cig⁻¹, without IcdP, DahA and BghiP),⁶ and Rustemeier et al. (630.8–912.4 ng cig⁻¹, without BbF and BkF),¹⁰ but lower than that measured in a commercial brand by Moir et al. (4910.7 ng cig⁻¹).⁷ Nevertheless, less discrepancy between this study and the literatures would occur when $\sum_{16}\text{PAHs}$ was normalized by TPM; the range of $\sum_{16}\text{PAHs}/\text{TPM}$ was 4.8–93.1 ng mg⁻¹ in our study and 57.3–104.7 ng mg⁻¹ in previous ones.^{7,10}

Carcinogenic Potency. BaP has been used as an indicator of total PAHs and their carcinogenicity.³¹ It is also the most frequently reported PAH in the literature on cigarette smoke studies.¹ BaP levels for most cigarette brands measured in this study were within the literature range (1.3–36.3, Table S2, Supporting Information), whereas relatively lower yield for Kent (1.1 ng cig⁻¹) and higher yields for JZ (41.6 ng cig⁻¹), LQ (37.4 ng cig⁻¹), and YZB (40.3 ng cig⁻¹) were observed. As BaP is only one of many carcinogenic PAHs in tobacco smoke, a more reasonable assessment of the carcinogenicity could be achieved by including more carcinogenic PAH species. As a consequence, the BaP equivalent concentration (BaP_{eq}) is typically calculated by using BaP as a reference congener and summing up the potencies of individual carcinogenic PAHs with their relative potency factors (RPFs, Table S2) obtained from USEPA.³² The carcinogenic potencies of individual PAHs and their contributions to the BaP_{eq} in the present study are listed in Table 3 in comparison with those in several Chinese,⁴ United States,^{5,16} Canadian,⁷ and European cigarette brands.³³

The BaP_{eq} values for the 13 commercial cigarettes ranged from 3.6 to 120.2 ng cig⁻¹, with an average of 71.7 ng cig⁻¹, comparable to the calculated values from Akpan et al.,⁴ Roemer et al.,⁵ Moir et al.,⁷ Ding et al.,¹⁶ and Lodovici et al.³³ BaP_{eq} levels seemed to be higher in flue-cured cigarettes (70.0–120.2 ng cig⁻¹) than in blended ones (3.6–52.5 ng cig⁻¹), except for similar levels in BS (60.8 ± 0.1 ng cig⁻¹) and WBL (66.1 ± 23.4 ng cig⁻¹). BaP and DahA were found to be the two largest contributors to BaP_{eq} with shares of 29.2–45.2% and 9.7–40.0%, respectively, similar to the results from Roemer et al. (29.2–30.8% and 38.1–41.2%)⁵ and Moir et al. (27.2% and 21.9%)⁷ (Table 3). However, DahA was the dominant contributor to BaP_{eq} in the 20 Chinese brands measured by Akpan et al. (68.1–85.7%),⁴ 9 United States brands reported by Ding et al. (52.9–66.0%),¹⁶ and 14 European brands measured by Lodovici et al. (40.8–81.1%, gas+particle phase).³³ It was noteworthy that the carcinogenic potency of the cigarette smoke was likely to be underestimated in this study because only a handful of PAHs were taken into account and to a lesser extent, only particle-phase concentrations were measured. For example, dibenzo[a,l]pyrene has a high RPF value of 30.³²

Table 3. Carcinogenic Potency of Individual PAHs and Their Contributions, ng cig⁻¹ (%), to the BaP_{eq}

| | this study | 20 Chinese brands ^a | 8 United States brands ^b | 9 United States brands ^c | 1 Canadian brand ^d | 14 European brands ^e |
|--------------------|----------------------|--------------------------------|-------------------------------------|-------------------------------------|-------------------------------|---------------------------------|
| Flu | 5.6–126.5 (5.7–12.6) | | | | 7.7 (14.7) | 1.0–5.6 (2.8–19.8) |
| CPcdP ^f | 0.5–12.1 (0.6–5.7) | | | | | |
| BaA | 2.0–56.4 (6.6–11.1) | 1.5–5.1 (1.5–5.6) | 2.4–5.4 (10.3–11.1) | 7.6–13.3 (10.0–16.0) | 6.1 (11.6) | 0.8–1.8 (1.8–7.3) |
| Chr | 2.5–64.4 (3.5–6.9) | 2.0–7.2 (1.8–8.3) | | | 3.9 (7.4) | 0.7–8.0 (2.8–18.6) |
| BbF | 0.5–17.7 (9.2–16.6) | 1.4–11.6 (1.2–11.6) | 4.6–9.1 (17.5–19.6) | 4.1–9.7 (7.0–12.5) | 8.6 (16.4) | 0.8–5.2 (1.8–11.3) |
| BkF | 0.6–17.1 (0.3–0.5) | 0.0–0.2 (0.0–0.1) | 0.1–0.1 (0.2–0.3) | 0.1–0.2 (0.1–0.2) | 0.1 (0.2) | 0.0–0.0 (0.0–0.1) |
| BaP | 1.1–41.6 (29.2–45.2) | 5.8–14.2 (5.1–12.4) | 7.2–14.5 (29.2–30.8) | 10.0–15.8 (13.8–19.0) | 14.3 (27.2) | 1.9–5.1 (5.6–17.7) |
| IcdP | 0.3–18.1 (0.5–1.5) | | 0.2–0.5 (0.8–1.0) | 0.5–0.8 (0.6–1.0) | 0.3 (0.6) | |
| DahA | 0.1–3.3 (9.7–40.0) | 32.0–189.0 (68.1–85.7) | 8.9–18.7 (38.1–41.2) | 36.0–62.0 (52.9–66.0) | 11.5 (21.9) | 9.0–72.0 (40.8–81.1) |
| BghiP | 0.1–6.9 (0.0–0.1) | | | | 0.0 (0.1) | 0.0–0.0 (0.0–0.1) |
| Sum | 3.6–120.2 (100) | 46.6–222.7 (100) | 23.3–48.3 (100) | 58.6–100.3 (100) | 52.6 (100) | 19.8–88.8 (100) |

^aRef 4. ^bRef 5. ^cRef 16. ^dRef 7. ^eRef 33. ^fCyclopenta[cd]pyrene.

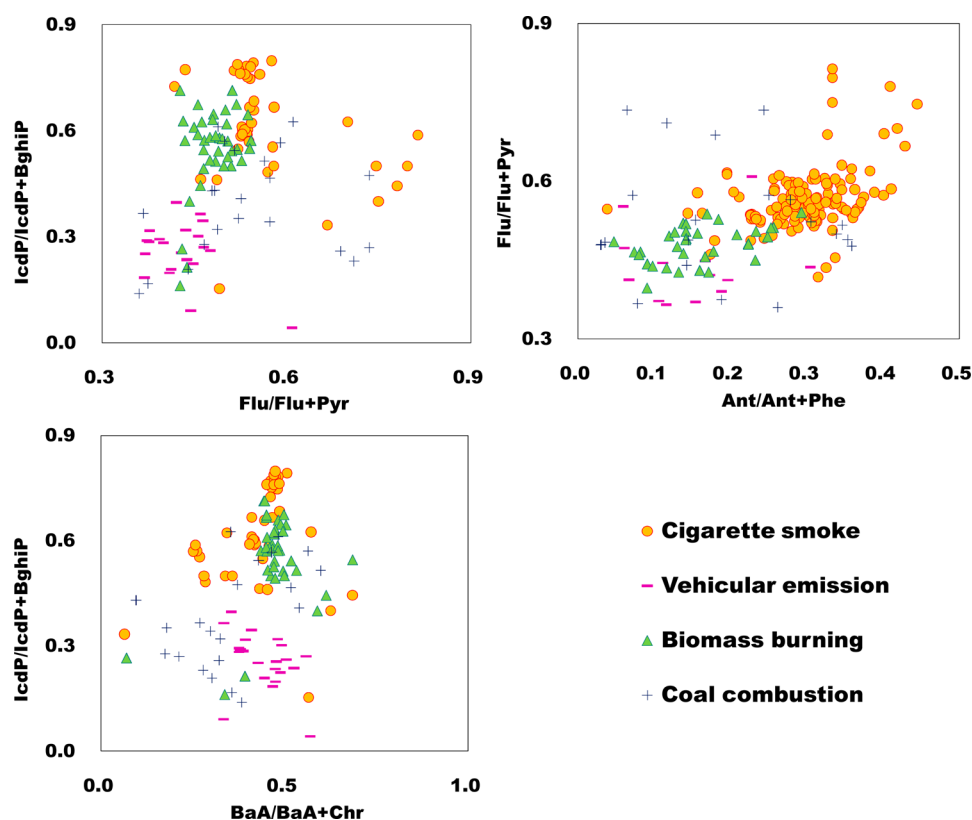


Figure 1. Cross plots for parent PAH ratios in cigarette smoke, vehicular emission, biomass burning, and coal combustion. Both values measured in this study and reported in the literature were used for cigarette smoke.

However, dibenzopyrenes were not included in the estimation of BaP_{eq} because the four isomers, dibenzo[a,e]pyrene, dibenzo[a,h]pyrene, dibenzo[a,i]pyrene, and dibenzo[a,l]pyrene were difficult to be cleanly separated by the GC column we used.

Discrimination of Cigarette Smoke from Other Sources Using Diagnostic Ratios. Parent PAHs. Ratios of several parent PAHs, such as Ant/(Ant+Phe), Flu/(Flu+Pyr), BaA/(BaA+Chr), IcdP/(IcdP+BghiP), and BaP/BghiP, have been widely used as diagnostic tools to trace possible emission sources.^{19,20} As vehicular exhausts (VE), biomass burning (BB), and coal combustion (CC) are documented as the three most common sources of PAHs,^{34–36} we compared these ratios derived from PAHs measured in particle-phase for cigarette smoke (including MS, sidestream smoke and ETS) with those for VE, BB, and CC (Table S3) to examine whether cigarette

smoke can be distinguished from other sources. Apparently, none of these ratios can differentiate the sources due to large variation for each source type and overlaps between different sources.

Figure 1 further compares the diagnostic ratios of typical PAH isomer pairs using cross plots. The dots of cigarette smoke exhibited remarkable overlap with those of BB and CC but were largely separate from that of VE, indicating the cross plots of PAH isomer pairs could possibly distinguish cigarette smoke from VE. However, caution is required for ambient data because the ratios may undergo change during transport due to differential partitioning, degradation and scavenging.³⁴ In addition, it is recognized that not all smokers smoke their cigarettes in the same way. Potential differences in smoking behavior, such as puff volume and frequency and blockage of cigarette ventilation holes, may give rise to changes in the

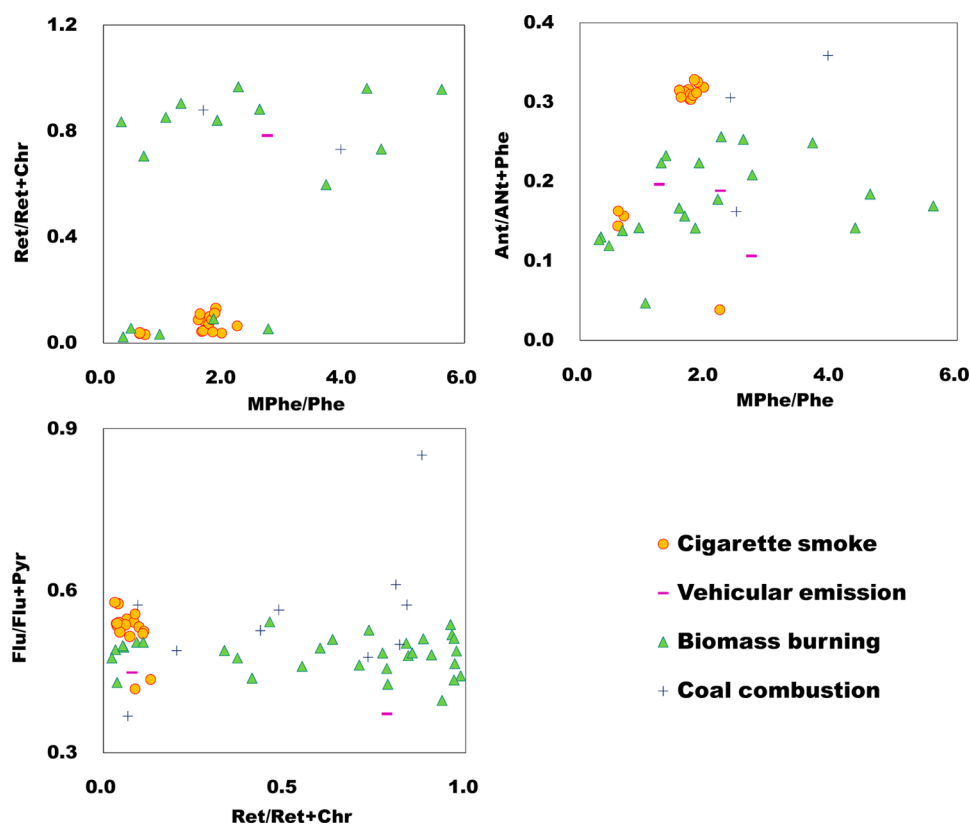


Figure 2. Cross plots for alkylated PAH ratios in cigarette smoke, vehicular emission, biomass burning, and coal combustion. Both values measured in this study and reported in the literature were used for cigarette smoke.

relative yields of tobacco combustion products.^{5,7,16} These factors will likely result in similar ratios for cigarette smoke and other sources.

Alkylated and S/O PAHs. Due to high structural variability, alkylated PAHs possess the potential to improve source discrimination. Relative concentrations of parent versus alkylated PAHs have been proposed to discriminate combustion from petrogenic emissions, with combustion characterized by a low proportion of alkyl-PAHs.³⁷ However, comparison of \sum parent versus \sum alkylated between this study and literature data is hindered by the fact that previous source profiles do not contain or only contain a few alkylated PAHs and their isomers. Measurements of all the parent, alkylated, and S/O PAHs are rarer. In a recent research, Schuster et al. suggested that the ratios \sum parent/ \sum alkylated ~ 0.2 and \sum parent/DBTs ~ 5 were generally representative of the Athabasca oil sands mining region,³⁸ and the ratios in another industrial region, Hamilton Harbour, Ontario, were 0.38 and 0.94, respectively. In this study, these ratios averaged 1.05 and 0.41, respectively, but the species and number of PAHs for calculating \sum parent, \sum alkylated, and DBTs had minor differences with those in Schuster et al.³⁸

Retene (Ret) has been used as a typical tracer for either wood combustion³⁹ or for degradation products of conifer resins.⁴⁰ DBT also acts as an indicator of combustion of sulfur-containing fuels such as coal and diesel.³⁹ However, both were detected in our MS sample and Ret was also detected in ETS,⁴¹ VE,^{42–44} and CC.^{45,46} Therefore, ratios containing these compounds rather than the absolute concentrations may be more suitable to distinguish sources. Ret/(Ret+Chr), Phe/(Phe+DBT), and DBT/(DBT+Ret) have been used to characterize the sources in environmental samples.^{22,37,47} Methylphenan-

threnes/phe (MPhe/Phe, MPhe includes 1-, 2-, 3-, and 9-MPhe) is also a frequently used indicator to differentiate between petrogenic and pyrogenic sources.^{24,48} Table S4 of the Supporting Information lists the ratios for different sources. Similar to parent PAHs, single ratios cannot satisfactorily distinguish cigarette smoke from other sources. Only a few ratios with S/O-containing PAHs were reported, such as Phe/(Phe+DBT) ~ 1.0 and DBT/(DBT+Ret) ~ 0.6 by Wingfors et al.⁴² for VE and Phe/(Phe+DBT) ~ 0.6 by Grimmer et al.⁴⁹ for CC. Unfortunately, these ratios fell into ranges for MS measured in this study. However, with the cross plots (Figure 2) involving more alkylated PAHs, the discrimination of cigarette smoke from CC seemed to be improved by the Ret/(Ret+Chr)-MPhe/Phe plot based on limited MPhe and Ret data in published source profiles.^{44,46,50–54} Anyway, future research is required when more source profiles containing alkylated and S/O PAHs are available.

Implications and Limitations. In this study, we obtained a much wider spectrum of PAHs in mainstream cigarette smoke for popular cigarette brands sold in China's markets. The doses of BaP and BaP_{eq} by smoking one cigarette of these brands averaged 26.6 and 75.9 ng, respectively; they were comparable or even higher than the daily doses for adults through inhalation exposure (31.2 and 42.3 ng, respectively), based on Nouwen equation,⁵⁵ in which the concentrations of particulate PAHs were taken from measured ambient values in urban Guangzhou on a day with highly polluted air.⁵⁶ Although we calculated BaP_{eq} to indicate carcinogenic potency, it must be stressed this simplified calculation only takes limited parent PAH species into consideration, and there are numerous alkylated and heteroatomic PAHs in cigarette smoke with health effects still unknown. The extended chemical speciation

of PAHs in cigarette smoke would not only help us understand air toxics released by smoking but also improve source attribution of PAHs. While cross plots of parent PAH pairs could only possibly distinguish cigarette smoke from vehicle exhausts, discrimination between cigarette smoke and coal combustion is enhanced with more alkylated PAHs involved. However, as cigarette smoke is in its nature a kind of biomass burning, it remains difficult to distinguish cigarettes smoke from other biomass burning sources. Compared to parent and alkylated PAHs, S/O PAHs and other heteroatomic PAHs were much less reported in published source profiles. To measure more heteroatomic PAHs in source profiles in future would benefit to exploring more fingerprints for better source attribution of PAHs occurring in indoor and outdoor environments. It is worth noting that this source attribution would be further complicated by atmospheric processes such as gas-particle partitioning and air mass aging.

■ ASSOCIATED CONTENT

■ Supporting Information

Tables outlining the characteristics of cigarettes smoked, yields of individual PAH and diagnostic ratios of parent, alkylated and S/O PAHs in cigarette smoke, and other sources. The Supporting Information is available free of charge on the ACS Publications website at DOI: 10.1021/acs.est.5b01108.

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■ Notes

The authors declare no competing financial interest.

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